

Reprinted from

# JOURNAL OF MICROBIOLOGICAL METHODS

---

Journal of Microbiological Methods, 25 (1996) 309–315

## Quantitative determination of *Escherichia coli* in water using CHROMagar® *E. coli*

Jose L. Alonso<sup>a,\*</sup>, Inmaculada Amoros<sup>a</sup>, Steven Chong<sup>b</sup>, Hemda Garelick<sup>b</sup>

<sup>a</sup>*Instituto de Hidrologia y Medio Natural, Universidad Politecnica, 46022 Valencia, Spain*

<sup>b</sup>*Urban Pollution Research Centre, School of Environmental Science and Engineering, Middlesex University, Bounds Green Road, London N11 2NQ, UK*

Received 18 April 1995; revised 15 November 1995; accepted 18 December 1995





## Quantitative determination of *Escherichia coli* in water using CHROMagar® *E. coli*

Jose L. Alonso<sup>a,\*</sup>, Inmaculada Amoros<sup>a</sup>, Steven Chong<sup>b</sup>, Hemda Garelick<sup>b</sup>

<sup>a</sup>Instituto de Hidrologia y Medio Natural, Universidad Politecnica, 46022 Valencia, Spain

<sup>b</sup>Urban Pollution Research Centre, School of Environmental Science and Engineering, Middlesex University, Bounds Green Road, London N11 2NQ, UK

Received 18 April 1995; revised 15 November 1995; accepted 18 December 1995

### Abstract

A new medium, CHROMagar® *E. coli* (CAEC), containing a combination of X-glucuronide and methyl-glucuronide for the detection of  $\beta$ -glucuronidase activity of *Escherichia coli* has been evaluated by the membrane filtration (MF) technique in fresh water samples. The CAEC agar was compared with conventional media, mFC agar and mLSB, for the enumeration of faecal coliforms. The variance analysis showed that CAEC was as sensitive as mFC agar and mLSB. A good correlation was found between *E. coli* versus those from faecal coliforms in the water sampling areas tested. Of 321 presumptively positive *E. coli* colonies (blue) and 154 presumptively negative *E. coli* colonies (white), only 8 (2.5%) false positive and 19 (12.4%) false negative colonies were found. Specificity of the CAEC agar in Spanish samples was temperature dependent, false negative *E. coli* colonies occurred less frequently at 37°C (2.3%) than at 44.5°C (18.8%). The results of this study indicate that CAEC agar is efficient for the enumeration of *E. coli* from a wide range of environmental freshwater samples.

**Keywords:** *Escherichia coli*; Faecal pollution; Freshwater; GUD; Membrane filtration

### 1. Introduction

The detection and enumeration of indicator organisms are of primary importance for the monitoring of sanitary and microbiological quality of water. Faecal coliforms have long used as indicators of faecal contamination in water and food. Among these, *Escherichia coli* is generally considered the most reliable, since its presence directly relates to faecal contamination with its implied threat of the presence of enteric disease agents [1].

A major limitation of the standard methods available for the enumeration of *E. coli* in water is the length of time required to complete the testing [2]. To overcome this problem, new enzymatic methods incorporating fluorogenic and chromogenic substrates have been developed for the presumptive detection of *E. coli*  $\beta$ -glucuronidase (GUD) (E.C. 3.2.1.31) is an enzyme that catalyzes the hydrolysis of  $\beta$ -D-glucopyranosiduronic (GLR) derivatives into their corresponding aglycons and D-glucuronic acid [3], 94–97% of *E. coli* were found to have positive GUD-activity [4], which is not expressed by other species of *Escherichia* [5]. Other members of the Enterobacteriaceae producing GUD activity include

\*Corresponding author. Tel.: +34 (6) 3877 090; Fax: +34(6) 3877 009; e-mail: jalonso@idhr.upr.es

17–29% of *Salmonella* sp. [6,7], 50% of *Shigella* [8] and a minority of *Yersinia* sp. [9]. Chromogenic and fluorogenic  $\beta$ -glucuronides incorporated in various liquid and solid media have served as indicators of GUD activity. The fluorogenic substrate 4-methylumbelliferyl- $\beta$ -D-glucuronide (MUG) has been used in the detection of *E. coli* in water by membrane filtration (MF) technique [10–15]. However, MUG analysis requires the use of fluorescence light rendering it less suitable for using in membrane filtration [14,16]. The chromogenic substrate *p*-nitrophenol- $\beta$ -D-glucuronide (PNPG) is also used to detect *E. coli*, but PNPG is unable to enumerate distinct *E. coli* colonies because the colour that develops during bacterial growth is not retained within the bacterial colony [17]. Recently, other chromogens, such as indoxyl derivatives, 5-bromo-4-chloro-3-indolyl- $\beta$ -D-glucuronide (X-Gluc) and indoxyl- $\beta$ -D-glucuronide (IBDG) have also been used to detect or enumerate *E. coli* in water [18–20]. These substrates release indoxyl (or halogenated indoxyl) which is rapidly oxidized to the insoluble indigo or its analogue, providing a very distinct differentiation of GUD-positive *E. coli* as blue colonies on agar plates or membranes [14].

This study was designed to evaluate CHROMagar® *E. coli*, containing a combination of X-glucuronide and methyl-glucuronide, as a new medium for the enumeration of *E. coli* by the membrane filtration technique in a variety of freshwater samples from Spain and England.

## 2. Materials and methods

### 2.1. Sampling

A series of freshwater samples from England and Spain were used to evaluate the ability of CAEC

agar to enumerate *E. coli*. Locations of sampling sites are indicated in Table 1. All samples were collected in sterile glass bottles, refrigerated and assayed within 3 h after collection.

### 2.2. Media

CHROMagar® *E. coli* (CAEC agar was obtained from CHROMagar (Paris, France). The composition of the powdered medium is in mg·l<sup>-1</sup>: peptone 5, yeast and meat extracts 3, sodium chloride 5, selective chromogenic mix 9, agar 15 (pH 6.0).

Membrane lauryl sulphate broth (mLSB) (Oxoid, Basingstoke, UK) and mFC broth (Difco, MI, USA) were prepared according to the manufacturer's instructions. In the case of mFC broth was solidified by adding 1.2% agar before boiling [21].

### 2.3. Media comparison

Dilutions of the water samples were made and duplicates of each dilution filtered through sterile 0.45- $\mu$ m porosity membranes (Millipore, Bedford, MA, USA) using the standard membrane filtration technique [21].

One of each set of duplicate membranes was placed onto a pre-prepared layer of CAEC agar in a 47 mm petri-dish. These were then incubated at 44.5°C for 24 h and all blue colonies (GUD+) were identified as presumptive *E. coli*. Non-blue colonies (GUD-) were counted as presumptive non-*E. coli* isolates.

For comparison, the second duplicate membrane of each duplicate pair was processed by a standard method for faecal coliforms. The membranes containing the English samples were placed onto sterile absorbent pads soaked with mLSB in 47 mm Petri-dishes. These were then incubated at 44.5°C for 24 h. All yellow colonies of a diameter greater than 1 mm

Table 1  
Locations of sampling sites in England and Spain

Sampling site	Pollution sources	Site code
Hadley woods, England (Pymmes Brook Source)	Wildfowl, fish, animals	S
Osidge Lane, England (Pymmes Brook-Waterfall Walk)	Sanitary wastewater, surface water sewers	OL
Waterfall Road, England (Pymmes Brook-Amos Park)	Combined sewerage, surface water sewers	WR
Telford Road, England (Bounds Green Brook)	Surface water sewers	TR
Irrigation channel, Spain	Sewage	S1
Vera irrigation channel, Spain (Alboraya village)	Sewage	S2

were counted as presumptive faecal coliforms [22]. The membranes containing the Spanish samples were layered onto mFC agar and were incubated at 44.5°C for 24 h. All blue colonies were counted as faecal coliforms [21].

#### 2.4. Colony verification and identification

Isolated colonies were picked from the most appropriate dilution of CAEC agar. For each sample, 2 to 5 GUD<sup>+</sup> (blue colonies) and 2 to 5 GUD<sup>-</sup> (white colonies) were randomly picked and subcultured on nutrient agar (Oxoid). Colonies picked were then confirmed as *E. coli* by gas production in EC broth (Oxoid) and indole production in tryptone water following incubation at 44.5°C for 24 h as recommended for the standard methods [21,22]. Isolates were further identified using the API 20E system (bioMerieux, Marcy-l'Etoile, France).

The results of specificity testing were expressed as two individual terms; the error introduced by false positive colonies and the error resulting from undetected target colonies.

#### 2.5. Accuracy of CAEC agar

The accuracy, defined as a measure of the closeness of observed values to a known true value, was determined using two reference capsules of *E. coli* WR1 (prepared at the SVM, RIVM, Bilthoven, The Netherlands). The mean bacterial numbers of reference capsules of the batches used in England and Spain were  $37.00 \pm 5.40$  cfu per 100 ml<sup>-1</sup> and  $35.70 \pm 9.28$  cfu per 100 ml<sup>-1</sup>, respectively. The content of each capsule was reconstituted in a tube with 10 ml of peptone saline solution (PS), ingredients (g·l<sup>-1</sup>): peptone 1, NaCl 8.5, as recommended by SVM. One ml of the resulting *E. coli* suspension was then added to 99 ml of PS and filtered through a 0.45- $\mu$ m membrane. The membranes were then layered onto CAEC agar and incubated at 44.5°C for 24 h as described for the environmental water samples. Five replicates per capsule were done.

#### 2.6. Regression analysis and statistical evaluation

Bacterial counts were logarithmically transformed prior to statistical treatment. Results were analyzed by linear regression to verify the linearity of the

relationship between *E. coli* and faecal indicators obtained with different media. A unifactorial variance analysis was performed on the means of the data. To examine the medium performance over a range of samples types and concentrations, the samples were grouped by sample site and by *E. coli* counts. All statistics were obtained using Statgraphics software.

### 3. Results and discussion

A series of 60 water samples was investigated in England and Spain for *E. coli* and faecal coliforms. The 12 water samples tested in Spain had *E. coli* counts ranging from  $5.8 \times 10^5$  cfu·100 ml<sup>-1</sup> to  $4 \times 10^6$  cfu·100 ml<sup>-1</sup> (Table 2). The 48 water samples analyzed in England had *E. coli* counts ranging from 160 cfu·100 ml<sup>-1</sup> to  $1.8 \times 10^6$  cfu·100 ml<sup>-1</sup> (Table 3). The data presented in Table 4 show very good correlation ( $P \leq 0.01$ ) between *E. coli* and faecal coliforms counts at the six sites tested. A total of 88 out of 98 faecal coliform colonies (89.8%) on mLSB were *E. coli*. Ten yellow colonies out of 98 on mLSB were not *E. coli*, with 8 colonies being identified as *Klebsiella pneumoniae* (8.2%), 1 colony as *Citrobacter freundii* (1%) and 1 colony as *Pasteurella* sp. (1%). Bordalo [23] investigating the recovery of faecal coliforms on mFC agar in various types of water, ranging from polluted seawater to unpolluted freshwater found that *E. coli* represent, on average 50.0–70.8% of the faecal coliform population. Faecal coliforms are widely used as indicators of faecal contamination of water [21]), although some authors questioned the usefulness of such

Table 2

Counts of *Escherichia coli* and faecal coliforms at sampling sites of Spain

Date (1994)	Site S1 (cfu·100 ml <sup>-1</sup> ) <sup>a</sup>		Site S2 (cfu·100 ml <sup>-1</sup> )	
	<i>E. coli</i>	F. coliforms <sup>b</sup>	<i>E. coli</i>	F. coliforms
7 March	$1.9 \times 10^6$	$2.6 \times 10^6$	$8.0 \times 10^5$	-
8 March	$2.3 \times 10^6$	$3.7 \times 10^6$	$5.8 \times 10^5$	$7.2 \times 10^5$
9 March	$1.8 \times 10^6$	$2.7 \times 10^6$	$7.8 \times 10^5$	$8.9 \times 10^5$
10 March	$1.1 \times 10^6$	$1.5 \times 10^6$	$1.2 \times 10^6$	$1.4 \times 10^6$
26 April	$4.0 \times 10^6$	$4.5 \times 10^6$	$2.8 \times 10^6$	$3.2 \times 10^6$
27 April	$3.8 \times 10^6$	$4.4 \times 10^6$	$2.0 \times 10^6$	$2.6 \times 10^6$

<sup>a</sup>Values are the mean of two replicates.

<sup>b</sup>Faecal coliforms.

Table 3  
Counts of *Escherichia coli* and faecal coliforms at sampling sites of England

Date (1994)	Site S (cfu 100 ml <sup>-1</sup> ) <sup>a</sup>		Site OL (cfu 100 ml <sup>-1</sup> )		Site WR (cfu 100 ml <sup>-1</sup> )		Site TR (cfu 100 ml <sup>-1</sup> )	
	<i>E. coli</i>	Faecal coliforms	<i>E. coli</i>	Faecal coliforms	<i>E. coli</i>	Faecal coliforms	<i>E. coli</i>	Faecal coliforms
3 May	820	700	1.2×10 <sup>4</sup>	1.9×10 <sup>4</sup>	1.4×10 <sup>4</sup>	2.3×10 <sup>4</sup>	1.5×10 <sup>5</sup>	1.4×10 <sup>4</sup>
10 May	2000	600	2.3×10 <sup>4</sup>	6.1×10 <sup>4</sup>	2.4×10 <sup>4</sup>	1.8×10 <sup>4</sup>	1.6×10 <sup>5</sup>	1.6×10 <sup>5</sup>
17 May	1800	2100	1.7×10 <sup>5</sup>	7.2×10 <sup>5</sup>	1.8×10 <sup>6</sup>	5.0×10 <sup>5</sup>	2.6×10 <sup>5</sup>	3.6×10 <sup>5</sup>
24 May	1900	1700	2.8×10 <sup>4</sup>	6.0×10 <sup>4</sup>	8.7×10 <sup>4</sup>	1.0×10 <sup>4</sup>	2.1×10 <sup>5</sup>	1.7×10 <sup>5</sup>
31 May	1900	1800	1.3×10 <sup>4</sup>	2.1×10 <sup>4</sup>	2.8×10 <sup>4</sup>	2.9×10 <sup>4</sup>	3.6×10 <sup>4</sup>	3.8×10 <sup>4</sup>
21 June	1500	1350	8.9×10 <sup>4</sup>	1.2×10 <sup>4</sup>	1.1×10 <sup>4</sup>	8.1×10 <sup>4</sup>	1.4×10 <sup>4</sup>	1.6×10 <sup>4</sup>
5 July	1000	1330	1.4×10 <sup>4</sup>	1.4×10 <sup>4</sup>	2.5×10 <sup>4</sup>	5.8×10 <sup>4</sup>	1.5×10 <sup>5</sup>	1.7×10 <sup>5</sup>
12 July	1100	1200	1.5×10 <sup>5</sup>	5.2×10 <sup>5</sup>	6.1×10 <sup>4</sup>	4.6×10 <sup>4</sup>	1.0×10 <sup>4</sup>	8.0×10 <sup>3</sup>
19 July	400	500	7.9×10 <sup>4</sup>	5.4×10 <sup>4</sup>	1.1×10 <sup>4</sup>	1.1×10 <sup>4</sup>	4.0×10 <sup>4</sup>	4.4×10 <sup>4</sup>
26 July	2500	1800	1.4×10 <sup>5</sup>	1.3×10 <sup>5</sup>	3.0×10 <sup>4</sup>	2.0×10 <sup>4</sup>	2.1×10 <sup>4</sup>	7.0×10 <sup>3</sup>
2 August	320	290	4.8×10 <sup>4</sup>	5.6×10 <sup>4</sup>	3.7×10 <sup>4</sup>	3.9×10 <sup>4</sup>	1.0×10 <sup>4</sup>	1.0×10 <sup>4</sup>
9 August	160	190	4.3×10 <sup>4</sup>	4.0×10 <sup>4</sup>	4.0×10 <sup>4</sup>	2.3×10 <sup>4</sup>	4.0×10 <sup>4</sup>	4.4×10 <sup>4</sup>

<sup>a</sup>Values are the mean of two replicates.

thermotolerant organisms as indicators [24,25]. *E. coli* is preferred because it is specific and most reliably reflects the faecal origin [13].

Counts of *E. coli* on CAEC agar from water samples in Spain and England were compared with faecal coliform counts obtained with the comparison media, mFC agar and mLSB. The overall ANOVA test revealed no significant differences with the *E. coli* counts on CAEC agar and faecal coliforms on mFC agar and mLSB. For purposes of analysis, the abundance range was fractionated into five intervals (101–1000; 1001–10 000; 10 001–100 000; 100 001–1 000 000; ≥1 000 000). Considering the results of the ANOVA test for each interval it was found a good level of agreement between *E. coli* and faecal coliform counts.

The mean bacterial numbers of *E. coli* WR 1 in capsules tested in Spain and in England were found to be 36.30±4.45 cfu·100 ml<sup>-1</sup> and 34.80±2.89

cfu·100 ml<sup>-1</sup>, respectively. Recovery of the *E. coli* reference strain was satisfactory on CAEC agar. An average of 101.6% in Spain and 94.1% in England was obtained. These results showed a good accuracy for the CAEC agar.

Apparent *E. coli* colonies on CAEC agar were blue and all other colonies were white. The identities of the two types of colonies are shown in Table 5. Of the 321 blue colonies 313 (95.7%) were confirmed as *E. coli* (Table 6). This confirmation rate is similar to that reported for other media with indoxyl derivatives [19,20]. Some investigators [20,26] found that media with indoxyl derivatives (IBDG, BCIG) proved the easiest to read and produced the least number of false positives and false negatives than media with MUG or PNPG. False positive results on CAEC agar were due to strains of *Klebsiella oxytoca*, *Enterobacter intermedium*, *Serratia fonticola*, *Pasteurella multocida* and *Yersinia intermedia*. Al-

Table 4  
Regression and correlation parameters from data obtained using chromogenic and reference media

Sample source	Media <sup>a</sup>	CC <sup>b</sup>	P	Intercept (a)	Slope (b)
S1	CAEC-mFC	0.963	≤0.01	-1.029	1.141
S2	CAEC-mFC	0.996	≤0.01	-0.052	0.996
S	CAEC-mLSB	0.893	≤0.01	0.081	0.987
OL	CAEC-mLSB	0.856	≤0.01	1.126	0.733
WR	CAEC-mLSB	0.930	≤0.01	-0.915	1.210
TR	CAEC-mLSB	0.971	≤0.01	0.794	0.839

<sup>a</sup> CAEC (*E. coli*): dependent variable; mFC and mLSB (Faecal coliforms): independent variable.

<sup>b</sup> Correlation coefficient.

Table 5  
Identification of colonies picked from CAEC agar using the API 20e system

Species isolated	GUD+ <sup>a</sup>	Species isolated	GUD– <sup>b</sup>
<i>Escherichia coli</i>	119	<i>Escherichia coli</i>	16
<i>Pasteurella multocida</i>	2	<i>Klebsiella pneumoniae</i>	48
<i>Enterobacter intermedium</i>	2	<i>Aeromonas hydrophila</i>	14
<i>Klebsiella oxytoca</i>	2	<i>Enterobacter cloacae</i>	9
<i>Serratia fonticola</i>	1	<i>Citrobacter freundii</i>	9
<i>Yersinia intermedia</i>	1	<i>Klebsiella oxytoca</i>	7
		<i>Enterobacter agglomerans</i>	4
		<i>Pasteurella multocida</i>	3
		<i>Salmonella</i> sp.	1
		<i>Salmonella paratyphi</i> A	1
		<i>Serratia rubidaea</i>	1
		<i>Pseudomonas putrefaciens</i>	1
		<i>Enterobacter sakazakii</i>	1
Total	127		115

<sup>a</sup> Number of blue colonies.

<sup>b</sup> Number of non-blue colonies.

though  $\beta$ -glucuronidase activity have been reported [5,18,27,28] in some strains of *Enterobacter agglomerans*, *Enterobacter amnigenus*, *Citrobacter freundii*, *Citrobacter amalonaticus*, *Escherichia vulneris*, *Yersinia kristensenii*, *Hafnia alvei*, *Aeromonas* sp. and *Acinetobacter* sp., their occurrence appears to be very infrequent [18]. The reason for the production of  $\beta$ -glucuronidase by these organisms is not known, but other investigators [19] have suggested that the reaction may be plasmid mediated.

There seems to be some controversy over the number of GUD false negative *E. coli* and therefore questions have been raised on its effectiveness, as GUD-based tests may underestimate the presence of *E. coli*. In the present investigation we have found *E. coli* false negative occurrences of 18.8% in Spain in comparison with 6% in England (Table 6). These differences may have been based on differences in

samples, and therefore the physiological condition of the bacteria could be responsible for the non-expression of enzyme activity as suggested by different investigators [2,29–31]. The higher frequency of false *E. coli* strains in more polluted sites of Spain could be accounted for on the basis of different survival rate. Coyne and Schuler [32] suggested that if GUD positive *E. coli* survive longer in the environment than GUD negative *E. coli*, a higher frequency of GUD negative would be expected in freshly contaminated sites. Some investigators [30,33–37] observed that part of the genetic sequences of the *uidA* gene, which encodes for the GUD enzyme, was present in most if not all *E. coli* isolates, regardless of the GUD-phenotype and the absence of GUD enzyme activity may be attributed to physiological state of the isolates. We have incubated the CAEC agar at 37°C in some spanish

Table 6  
Verification of colonies observed on CAEC agar<sup>a</sup>

Sampling country	T <sup>b</sup>	No. of blue colonies (GUD <sup>+</sup> )		No. of non-blue colonies (GUD <sup>–</sup> )	
		<i>E. coli</i>	non- <i>E. coli</i>	<i>E. coli</i>	non- <i>E. coli</i>
Spain	44°C	218 (98.2)	4 (1.8)	19 (18.8)	82 (81.2)
	37°C	–	–	3 (2.3)	129 (97.7)
England	44°C	95 (95.9)	4 (4.1)	6 (6.0)	94 (94.0)
Total	44°C	313 (97.5)	8 (2.5)	25 (12.4)	176 (87.6)

<sup>a</sup> The percentage of isolates which were picked at random, tested and confirmed as *E. coli* or non-*E. coli* is in parentheses.

<sup>b</sup> Incubation temperature of CAEC agar.

samples to determine whether the expression of GUD production was temperature dependent. *E. coli* false negative colonies occurred less frequently at 37°C (2.3%) than at the higher temperature, 44.5°C (18.8%) (Table 6). From the data, we can assume that specificity of the CAEC agar in Spanish samples could be related to the physiological condition of *E. coli* isolates.

The results of this study indicate that the CAEC agar with the membrane filtration technique is satisfactory for the selective enumeration of *E. coli* in freshwater samples with different degree of contamination. Although, it is recommended that before its use elsewhere, a minimal number of samples should be examined locally to determine whether the specificity is affected by local environments.

### Acknowledgments

Thanks are due to A. Rambach for critical reading of the manuscript and C. Sabater for the statistical analysis of the data.

### References

- [1] Rice, E.W., Allen, M.J., Brenner, D.J. and Edberg, S.C. (1991) Assay for  $\beta$ -glucuronidase in species of the genus *Escherichia* and its applications for drinking-water analysis. *Appl. Environ. Microbiol.* 57, 592–593.
- [2] Balebona, M.C., Moriñigo, M.A., Cornax, R., Borrego, J.J., Torregrosa, V.M. and Gauthier, M.J. (1990) Modified most-probable-number technique for the specific determination of *Escherichia coli* from environmental samples using a fluorogenic method. *J. Microbiol. Methods.* 12, 235–245.
- [3] Manafi, M., Kneifel, W. and Bascomb, S. (1991) Fluorogenic and chromogenic substrates used in bacterial diagnostics. *Microbiol. Rev.* 55, 335–348.
- [4] Frampton, E.W. and Restaino, L. (1993) Methods for *Escherichia coli* identification in food, water and clinical samples based on beta-glucuronidase detection. *J. Appl. Bacteriol.* 74, 223–233.
- [5] Kämpfer, P., Rauhoff, D. and Dott, W. (1991) Glycosidase profiles of members of the family *Enterobacteriaceae*. *J. Clin. Microbiol.* 29, 2877–2879.
- [6] Feng, P.C.S. and Hartman, P.A. (1982) Fluorogenic assays for immediate confirmation of *Escherichia coli* from sewage and surface waters. *Appl. Environ. Microbiol.* 53, 1246–1250.
- [7] LeMinor, L. (1979) Tetrathionate-reductase,  $\beta$ -glucuronidase and ONPG-test in the genus *Salmonella*. *Zbl. Bacteriol. Hyg. Abt. Orig. A* 243, 321–325.
- [8] Kilian, N. and Bülow, P. (1976) Rapid diagnosis of *Enterobacteriaceae*. *Acta Pathol. Microbiol. Scand. Sect. B* 84, 245–251.
- [9] Petzel, J.P. and Hartman, P.A. (1985) Monensin-based medium for determination of total gram-negative bacteria and *Escherichia coli*. *Appl. Environ. Microbiol.* 49, 925–933.
- [10] Freier, T.A. and Hartman, P.A. (1987) Improved membrane filtration media for enumeration of total coliforms and *Escherichia coli* from sewage and surface waters. *Appl. Environ. Microbiol.* 53, 1246–1250.
- [11] Augoustinos, M.T., Grabow, N.A. and Kfir, R. (1993) An improved membrane filtration method for enumeration of faecal coliforms and *E. coli* by a fluorogenic  $\beta$ -glucuronidase assay. *Water Sci. Tech.* 27, 267–270.
- [12] Lewis, C.M. and Mak, J.L. (1989) Comparison of membrane filtration and autoanalysis Colilert presence-absence techniques for analysis of total coliforms and *Escherichia coli* in drinking water samples. *Appl. Environ. Microbiol.* 55, 3091–3094.
- [13] Shadix, L.C., Dunnigan, M.E. and Rice, E.W. (1993) Detection of *Escherichia coli* by the nutrient agar plus 4-methylumbelliferyl- $\beta$ -D-glucuronide (MUG) membrane filter method. *Can. J. Microbiol.* 39, 1066–1070.
- [14] Schets, F.M. and Havelaar, A.H. (1991) Comparison of indole production and  $\beta$ -glucuronidase activity for the detection of *Escherichia coli* in a membrane filtration method. *Lett. Appl. Microbiol.* 13, 272–274.
- [15] Cenci, G., De Bartolomeo, A. and Caldini, G. (1993) Comparison of fluorogenic and conventional membrane filter media for enumerating coliform bacteria. *Microbios* 76, 47–54.
- [16] Farber, J.M. (1986) Potential use of membrane filter and a fluorogenic reagent-based solid medium for the enumeration of *Escherichia coli* in foods. *Can. Inst. Food Sci. Technol.* 19, 34–37.
- [17] Ley, A.N., Bowers, R.J. and Wolfe, S. (1988) Indoxyl- $\beta$ -D-glucuronide, a novel chromogenic reagent for the specific detection and enumeration of *Escherichia coli* in environmental samples. *Can. J. Microbiol.* 34, 690–693.
- [18] Sartory, D.P. and Howard, L. (1992) A medium detecting  $\beta$ -glucuronidase for the simultaneous membrane filtration enumeration of *Escherichia coli* and coliforms from drinking water. *Lett. Appl. Microbiol.* 15, 273–276.
- [19] Brenner, K.P., Rankin, C.C., Roybal, Y.R., Stelma Jr., G.N., Scarpino, P.V. and Dufour, A.P. (1993) New medium for the simultaneous detection of total coliforms and *Escherichia coli* in water. *Appl. Environ. Microbiol.* 59, 3534–3544.
- [20] Haines, J.R., Covert, T.C. and Rankin, C.C. (1993) Evaluation of indoxyl- $\beta$ -D-glucuronide as a chromogen in media specific for *Escherichia coli*. *Appl. Environ. Microbiol.* 59, 2758–2759.
- [21] APHA, AWWA and WPCF (1989) Standard Methods for Examination of Water and Wastewater, 17th edition, American Public Health Association, Washington, DC.

- [22] Department of the Environment (1983) The Bacteriological Examination of Drinking Water Supplies. Reports on Public Health and Medical Subjects No. 71. HMSO, London.
- [23] Bordalo, A.A. (1994) Fecal coliform recovery in two standard media long an estuarine gradient. *Water Res.* 28, 2331–2334.
- [24] Dufour, A.P. (1977) *Escherichia coli*: the fecal coliform. In: Bacterial Indicators/Health Hazards Associated with Water (Hoadley, A. and Dutka, B.J., eds.), pp. 48–58, American Society for Testing Materials, Philadelphia, PA.
- [25] Evison, L.M. (1988) Comparative studies on the survival of indicator organisms and pathogens in fresh and sea water. *Water Sci. Tech.* 20, 309–315.
- [26] Odgen, I.D. and Watt, A.J. (1991) An evaluation of fluorogenic and chromogenic assays for the direct enumeration of *Escherichia coli*. *Lett. Appl. Microbiol.* 13, 212–215.
- [27] Watkins, W.D., Rippey, S.R., Clavet, C.R., Kelley-Reitz, D.J. and Burkhardt, W. (1988) Novel compound for identifying *Escherichia coli*. *Appl. Environ. Microbiol.* 54, 1874–1875.
- [28] Heizmann, W., Döller, P.C., Gütbrod, B. and Werner, H. (1988) Rapid identification of *Escherichia coli* by fluorocult media and positive indole reaction. *J. Clin. Microbiol.* 26, 2682–2684.
- [29] Sarhan, H.R. and Foster, H.A. (1991) A rapid fluorogenic method for the detection of *Escherichia coli* by the production of  $\beta$ -glucuronidase. *J. Appl. Bacteriol.* 70, 394–400.
- [30] Martins, M.T., Rivera, I.G., Clark, D.L., Stewart, M.H., Wolfe, R.L. and Olson, B.H. (1993) Distribution of *uidA* gene sequences in *Escherichia coli* isolates in water sources and comparison with the expression of  $\beta$ -glucuronidase activity in 4-methylumbelliferyl- $\beta$ -D-glucuronide media. *Appl. Environ. Microbiol.* 59, 2271–2276.
- [31] Clark, D.L., Milner, B.B., Stewart, M.H., Wolfe, R.L. and Olson, B.H. (1991) Comparative study of commercial 4-methyl-umbelliferyl- $\beta$ -D-glucuronide preparations with the Standard Methods membrane filtration fecal coliform test for the detection of *Escherichia coli* in water samples. *Appl. Environ. Microbiol.* 57, 1528–1534.
- [32] Coyne, M.S. and Schuler, J.C. (1994) Frequency of MUG negative *Escherichia coli* in Kentucky groundwater samples. *J. Environ. Qual.* 23, 126–129.
- [33] Feng, P., Lum, R. and Chang, G.W. (1991) Identification of *uidA* gene sequences in  $\beta$ -D-glucuronide-negative *Escherichia coli*. *Appl. Environ. Microbiol.* 57, 320–323.
- [34] Bej, A.K., McCarty, S.C. and Atlas, R.M. (1991) Detection of coliform bacteria and *Escherichia coli* by multiplex-polymerase chain reaction: comparison with defined substrate and plating methods for water quality monitoring. *Appl. Environ. Microbiol.* 57, 2429–2432.
- [35] Green, D.H., Lewis, G.D., Rodtong, S. and Loutit, M.W. (1991) Detection of faecal pollution in water by an *Escherichia coli uidA* gene probe. *J. Microbiol. Methods* 13, 207–214.
- [36] Palmer, C.J., Tsai, Y., Lang, A. and Sangermano, L.R. (1993) Evaluation of Colilert-marine water for detection of total coliforms and *Escherichia coli* in the marine environment. *Appl. Environ. Microbiol.* 59, 786–790.
- [37] Cleuziat, P. and Robert-Baudoy, J. (1990) Specific detection of *Escherichia coli* and *Shigella* species using fragments of genes coding for  $\beta$ -glucuronidase. *FEMS Microbiol. Lett.* 72, 315–322.